

6R: A Look at Resilient Vegetation and Growing Media

GROWING RESILIENCE: LONG-TERM PLANT DYNAMICS AND GREEN ROOF PERFORMANCE

Stephanie Carlisle, Max Piana

KieranTimberlake

Abstract

The benefits of green roofs are derived from their existence as functional, living ecosystems. While the architectural elements of a green roof assembly can be thought of as fixed, the biological components of a roof, its vegetation and growing media, are dynamic. This study presents findings from two mature intensive green roofs surveyed six to seven years after installation in Ithaca, New York. While the two green roofs are located on buildings of similar design in close proximity, the roofs vary in initial planting, detailing, and biophysical and microclimatic conditions. Vegetative surveying was utilized to explore changes in plant community structure and establish spatially explicit performance indicators, including species richness, cover, and biodiversity. Additionally, this data was paired with modeling of solar radiation exposure, exploring how site context influences community dynamics.

Introduction

Green roofs are complex dynamic systems capable of performing a wide range of ecosystem services—managing flows of water, energy, waste, nutrients and organisms—and demonstrating value over the lifetimes of the buildings on which they grow (26,39,43). As designers, landscape architects, engineers, building owners, and policy makers have become more aware of both the benefits of green roofs and the variability of project-specific returns, there has been an increased interest in the ability to assure the overall quality and performance of green roofs through the optimization of green roof assemblies and detailing (including plant selection, media mix design, trays, fabrics, drainage layers, irrigation techniques) (1,17,32).



For over a decade, researchers have approached green roofs as a "horticultural or engineering challenge" (10,33), and studies of green roof vegetation have primarily focused on short-term, controlled experiments evaluating single or limited species assemblies in replicated tray systems and small-scale deployments (13,21,29). While such research is empirically valuable, it is difficult to transfer the results of this research to practice. Green roofs installed on actual buildings bear little resemblance to the highly controlled and extensively maintained small-scale plots utilized by researchers, many of which are abandoned before plant communities are fully established (34).

While green roofs are carefully engineered to function as high-performance infrastructural elements, they are also living systems. No matter how well a green roof is initially designed and specified, all living systems grow and change over time. Over the life cycle of a building, plants installed on a green roof become established, mature, die, and regenerate as the roof is exposed to disturbances. The environmental context and conditions on green roofs are also spatially and temporally variable. Solar radiation is rarely evenly distributed over a site. Surrounding built context changes as neighborhood and landscape elements grow. Climate varies from year to year. Even growing media and roof drainage are not perfectly stable over the life of a green roof installation. If such changes in green roof context and composition are inevitable, what is their effect on performance? Adopting an ecological perspective can provide insight into growth dynamics of these living systems over time.

Even as an ecological perspective on green roofs has gained traction within the research community, our understanding of plant composition, functional diversity, and the long-term dynamics of green roof vegetation and resultant system function and performance is limited (7,33). Greenhouse studies reveal that plant composition affects different performance attributes, such as stormwater management (12,46), in controlled environments. However, efforts to accurately predict the performance of green roofs over a building's lifetime are challenged by a lack of long-term data on real buildings, as well as a shortage of research on the long-term dynamics of green roof plant communities (7).

Using the case study of two mature green roofs, each with more than five years of undisturbed growth, this study presents a unique opportunity to examine changes to vegetation communities, and the resilience of green roof communities. This research is part of a larger green roof monitoring agenda, from which a methodology has been established and applied to six roofs, ranging from two to ten years in age. The two roofs presented in this paper have been selected for comparison given their similarity in location and scale, but difference in initial plant composition. Within this context of understanding, the objectives of this study are threefold: 1) to describe the growth trajectories of two green roofs from establishment to maturity, 2) to establish a methodology for evaluating vegetative change and performance attributes, and 3) to discuss performance and resilience as they relate to species dynamics and the relationship between initial planting and emergent species. Ultimately, such an understanding of green roofs as adaptive, ecological systems will aid in predicting performance over time, and better inform the design and maintenance of resilient, high-performance roofscapes.



Study Site

The two green roofs examined in this study are located on the Cornell University campus in Ithaca, New York, atop separate dormitory buildings: the Alice H. Cook House (House 1) and the Carl L. Becker House (House 2). The two roofs were constructed one year apart, in 2005 and 2006 respectively, while census of green roof vegetation was completed in August 2012.

Ithaca, located in upstate New York, experiences a moderate continental climate, defined by warm, humid summers (average temperature July=20.4°C) and cold winters (average temperature January=-5.2°C) (31). The study area typically experiences 163-183 "freeze-free" days annually (8) and is located in a region that borders Plant Hardiness Zones 5b and 6a (41). Annual precipitation for the region is 93.98 cm and is distributed evenly over the year. During the summer of 2012, Ithaca experienced drought and historic temperature highs (31), conditions that are expected to have significantly impacted this study.

The House 1 green roof is an intensive green roof system covering 329 m² (47% of total roof area) divided by elevated skylights running east to west, which effectively establish four separate bays of vegetation. Originally planted with 16 species, the roof was designed to include a warm-season meadow mix of grasses and herbaceous forbs. The green roof has a total depth of 24.13 cm (including drainage layer) with approximately 20.32 cm of growing media, above a combination of Fabrene fabric, PVC membrane, and tapered rigid insulation. Two four-story dormitories to the north and south of the roof effectively "canyonize" the roof.

The House 2 green roof is physically similar to House 1, with vegetation representing 418 m² (50% of total roof area) divided by elevated skylights into four separate bays. Unlike House 1, this roof is an extensive green roof, originally designed to include only five species and to be primarily defined by three succulent plant species. The green roof features 12.70 cm of growing media on top of a sheet water retention layer and filter fabric. Adjacent buildings similarly canyonize the roof, reducing solar radiation and increasing shading from skylights. Unlike House 1, the eastern edge features a knee wall that acts as a significant wind block.



Figure 1 Research site. House 1 (left) is an intensive roof, and House 2 (right) is an extensive roof, both situated at Cornell University in Ithaca, New York. Campus image from Google Earth.



Methods

To assess green roof plant dynamics and vegetative performance, this study utilizes a spatially explicit survey methodology of plant species and roof conditions. The study method is based on the comparison of annual surveys to the original green roof installation, creating snapshots of the vegetative composition and associated performance characteristics.

Data Collection

The field methodology utilized in this study is based on the Relevé Method (Table 2), the most accepted method for vegetative surveys in Europe and the United States (22,37,44). In the field, each roof was segmented into 2 m² quadrants, and location and identity of plants were recorded through field diagrams of species footprints per quadrant. Quantitative analysis was captured and analyzed according to cover classes designated by the Braun-Blanquest cover/abundance scale (30,45), allowing the survey to accurately document growth and coverage by a mixture of plant types, including easily identifiable individuals and clonal species. Percent cover was recorded by calculating relative area occupied by the vertical projection of all aerial parts of plants, expressed as a percentage of the surface area of the sample plot at time of survey. Species names were identified and recorded for each plant species making up at least 5 percent of the cover in any one quadrant.

Data Analysis

Data from the surveys was assembled to identify roof composition and establish basic vegetation performance metrics. Analysis considered percent area of vegetative cover, plant type and species presence, distribution and sociability, species richness, and biodiversity. Vegetative cover, species richness, biodiversity, and sociability were computed across the entire roof as well as for each plot, allowing for a compiled plot-based view of the roof. Vegetative cover and sociability were considered at a species-specific level.

Analysis of census results was primarily focused at a species-specific level, but also considered plant type and family. *Plant type distribution* for each roof was calculated for the original plant selection and the census results. Species were categorized into one of eight plant types, including Grasses, Herbaceous Forbs, Shrubs, Trees, Vines, Succulents, and Bryophytes, according to USDA Plant Database plant type designations. Plant life forms represent different resource use patterns, adaptations to the external environment, and life history strategies; such plant forms can be considered a coarse proxy for functional diversity within a plant community (25), a method previously used in green roof studies (27).

Sociability is a plot scale measure of a species' tendency to exclude other species and form large groups or patches, or conversely to integrate with neighboring species through distributed or less dense growth patterns (18). Sociability scores represent the percent coverage of a species when present in a plot, relative to other species within that same plot, such that the higher the sociability score, the more likely that species is to occur in homogenous patches. *Species diversity* was calculated using the Shannon Index (35) and converted to true diversity (TD= $e^{Shannon Index}$) (19). While precedent is limited, the Shannon Index has been used previously by researchers (2,3,20) as an indicator of relative biodiversity and recognized for its ability to allow for a summary and comparison of biodiversity over time or across multiple roofs (2).



In order to assess microclimatic conditions on the roof related to solar access, a 3-dimensional context model of the building site was created in Rhinoceros3D and used to calculate the shading conditions on a plot basis for each green roof. A custom Grasshopper plugin was developed to calculate the average hours of direct solar radiation per day across peak growing season (May 1-October 1) by determining on an hourly basis whether the sample plot is occluded by context geometry (adjacent buildings or skylights) from direct sunlight. The Grasshopper plugin utilizes NOAA's solar angle calculator to determine solar angles appropriate for Ithaca, New York (42). Model results were averaged within each census plot and linearly regressed with plot-level vegetative cover and biodiversity scores using R software (38). A general comparison of House 1 and House 2 solar radiation was calculated with a one-way ANOVA test. Precedent for solar radiation studies on green roofs is limited (15) and often based on coarse regional data or even gualitative observation (9.24). Furthermore, while the consideration of variable light conditions in selection plants has emerged, the relationship between solar microclimates and compositional changes in rooftop vegetation have not been explored with rigor. The methodology utilized in this paper matches a resolution of analysis commonly employed in the design of other building elements, which has the potential to be extended to green roofs.

Data Interpretation

Data interpretation was achieved through traditional quantitative analysis and visually based diagramming of spatial relationships, at a species-specific or plot-level analysis. Survey notes were compiled and transcribed to produce a general map of the vegetative composition, as well as a plot-based map of the vegetation performance metrics (coverage, richness, diversity). The combination of quantitative and graphic processing of survey results allows one to observe plant community relationships over time and focus on specific zones or growth patterns that may not be apparent from numerical outputs alone. The ability to isolate discrete populations of plants—for example, by species, function group, or point of emergence—also provides a unique opportunity to examine changes in the roof system and begin a dialogue that considers resiliency and performance, from design intent to present day conditions.

Results

In the seven years since initial planting, House 1 has transitioned from a mixed meadow roof to a single-species dominated roof system, in which the designed planting zones have been obscured by the colonization of *Schizachyrium scoparium*. While the 2012 census results found the roof to include 39 distinct species, including 14 of the original 16 species, *S. scoparium* presently represents more than 55% of total vegetative coverage on the roof. Of the other originally planted species, none were found to contribute more than 5% to total roof vegetation. The second most abundant species was an emergent, *Melilotus officinalis*, which represented 13.76% of all vegetative cover at time of survey. Collectively, ruderal or emergent species represented 31.25% of all vegetative cover. As such, while the species richness of the roof has more than doubled, species biodiversity calculations reveal that, as a system, the House 1 green roof was less diverse in 2012 (True Diversity (TD) = 6.74) then at time of planting in 2005 (TD=8.67). Still, despite these dramatic changes in species representation, the roof maintains nearly full coverage (93%).



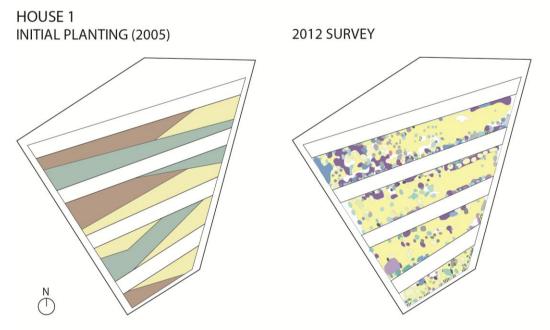


Figure 2 House 1 was initially planted with 16 species in three distinct zones (evergreen cover, low meadow, and high meadow). At the time of survey, 39 species were identified on the roof, including 14 of the original species.

Conversely, the dynamics of the House 2 green roof vegetation over the past seven years are defined by the resilience and persistence of designed planting zones, which have effectively supported increased diversity and ecological complexity across the roof. Originally planted with only five species (including three species of sedum, one warm season grass, and one herbaceous forb), the roof now features 65 distinct species, representing 30 plant families. The roof census reveals that sedum plantings still dominate the system and have maintained coverage in designated planting zones, but they have allowed for the integration of a variety of ruderal species, including forbs, shrubs, and trees, throughout the roof.

The House 2 green roof exhibits nearly full coverage (95%), with multi-strata communities appearing as succulents that occupy area beneath emergent forbs and trees. Additionally, House 2 species diversity increased dramatically from original planting (TD=3.67) to 2012

(TD=14.80). Species diversity on House 2, while more evenly distributed across the roof, was found to be greater on the southernmost bay (Bay 1=16.78 P value). Upon further investigation, this trend appears to be the result of microhabitat conditions associated in part with solar exposure and increase in shelter from the adjacent building.

House 2	True Diversity	Species Presence	Percent Coverage
Bay 1	16.78	31	99.68%
Bay 2	4.20	28	96.11%
Bay 3	4.54	30	91.03%
Bay 4	5.54	24	86.88%

Note: Bays are numbered 1-4 from south to north.



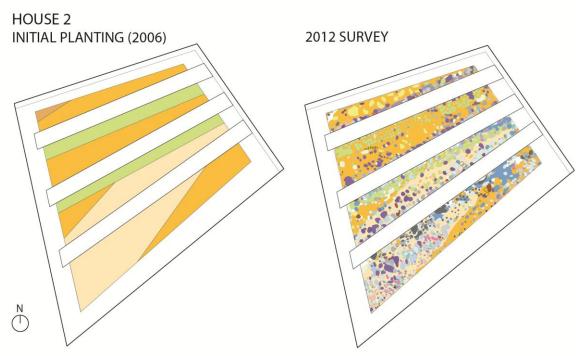


Figure 3 House 2 was initially planted with five species (four succulents and one herbaceous forb). At the time of survey, 65 species were identified on the roof, including all five original species.

Building	Year	Percent Coverage	Species Richness	Plant Families	True Diversity (e ^{Shannon})
House 1	2005	NA	16	7	8.41
nouse i	2012	93%	39	17	6.98
House 2	2006	NA	5	3	3.67
nouse 2	2012	95%	65	30	14.89

Comparisons between the mapping of vegetative composition (species level) and vegetation performance metrics (percent coverage, species presence, and diversity) portray both roofs as diverse, heterogeneous landscapes. At a plot level, species richness and biodiversity were variable across both roofs (Fig. 4) with concentrated pockets of high diversity and richness. Floristic relationships and performance indicators may be the result of microhabitat heterogeneity related in part to the uneven distribution of resources (light, nutrients, and moisture) and the contribution of context features (adjacent buildings, skylights, and site walls) in mitigating disturbances by acting as wind breaks and shading devices.

Mapping of performance measures pulls these features into focus, with areas of highest biodiversity, species richness, and percent cover occurring in bays closest to an adjacent building face (to the north and south of House 1, and to the south and west of House 2). At a smaller scale, the shading and thermal protection offered by the elevated skylights consistently produce a microhabitat zone capable of supporting a unique collection of emergent trees and herbaceous forbs not found in other areas of the roof.



Figure 4 Vegetative performance measures expressed across sample plots, summer 2012 survey.

Solar Analysis

Microclimate analysis revealed that solar radiation for both houses is variable, as a result of neighboring buildings and elevated skylights. Comparatively, House 1 (Mean=10.18) receives more hours of sunlight than House 2 (Mean=7.48), F(1,209)=90.22, p=0.000. Qualitatively, it is readily apparent that in areas of greatest solar exposure (typically more than 10 hours per day), vegetative coverage and species diversity decrease for both House 1 and House 2. Given that previous studies have related moisture stress and solar exposure to plant performance broadly and specifically (10,15,28), it is not surprising that areas with greater solar access exhibited less plant diversity and lower vegetative coverage.

House 2 regression analyses found increased solar exposure to negatively impact species diversity (R^2 =20.79, p<0.000) and vegetative cover (R^2 =26.30, p<0.000) at the plot scale. Similar analyses of House 1 did not prove to be statistically significant; this is, however, explained by the resolution of plot data from census grid (2 m²). Solar analysis was run at 30 cm grids across each roof, the results of which capture microclimate factors that correspond with qualitative field observations. Such observations include increased species diversity and presence along each skylight and at the roof's edges. Future efforts should be extended to match the resolution of performance metrics with the finer grid analysis.

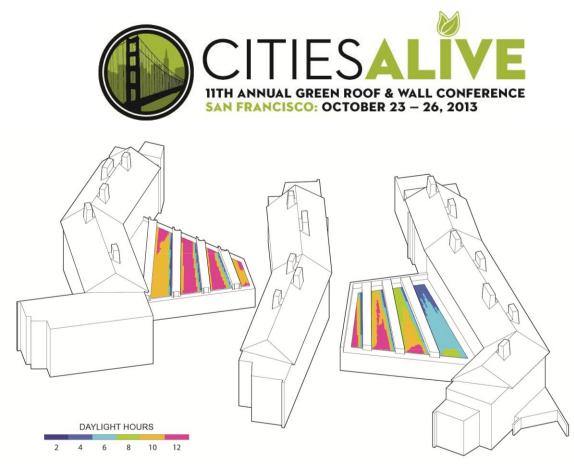


Figure 5 Solar access analysis for House 1 (left) and House 2 (right).

Discussion

Over the past seven years, the green roofs on House 1 and 2 have grown with minimal human intervention, allowing for patterns of dynamic change to occur. As a result, this study has been able to capture two distinct and divergent growth trajectories or narratives of mature green roof systems. Generally, while both green roofs displayed high percent cover and good vegetative health at time of survey, House 2 has increased in plant diversity and system complexity since its initial installation while House 1 has decreased in system complexity, transitioning into a single-species dominated system.

The observed changes on both roofs are reflective of a combination of static and dynamic environmental conditions and floristic associations. From an ecological perspective, changes to a plant community operate over temporal and spatial scales. Temporally, one may observe phenological changes during the growing season or over a year, fluctuations or cycles over multiple years, or in the long-term, successional changes in plant composition. Throughout these temporal phases, spatial change may occur at the individual plant-plant level, between plant communities, and most broadly, at the landscape scale (11). Ultimately, both spatial and temporal changes are driven by lifecycles of the organism and population, as well as the biophysical factors and constraints at the level of the individual to the landscape (11).

The green roofs studied in this paper are designed landscapes, installed as part of constructed and mediated ecosystems with a diversity of initial conditions, including an uneven distribution of plant species in planting zones. Over the last seven years, plant communities on the roofs have responded to this heterogeneous landscape—collectively becoming tuned to the stable microclimates of the site (such as solar access and wind patterns) while responding to irregular



disturbances such as drought, disease, and competing emergent species. How do these changes relate to roof performance? Is the presence of emergent species on these roofs the sign of benign neglect, or a symbol of increased biodiversity and community complexity?

Biodiversity, as suggested by Orbendoffer et al., is significant in determining a green roof system's resilience to disturbances as well as its efficiency of resource consumption (33). Recent research supports this perspective, relating not only plant diversity but specific combinations of plant species to greater achieved performance of these systems (27). The current state of vegetation on House 2 is an example of such a combination, in which a small number of originally planted sedum species has been enriched by 50 ruderal species, representing 30 plant families cohabitating to create a rich multi-strata assembly that proved extremely resilient to the drought conditions at time of survey. This increase in diversity and system complexity on House 2 is likely the result of both interspecies facilitation by the sedum species and microclimate conditions defined by site context. Like some desert nurse species, which create microclimates where soil is cooled and moisture increased (14,39), green roof research has found sedum to facilitate growth and survival of neighboring plant species during periods of environmental stress (4,5).

Alternatively, while House 1 exhibits nearly full coverage, *Schizachyrium scoparium* has successfully outcompeted and colonized much of the roof, moving the system towards a state of reduced diversity and floristic complexity. This decrease in biodiversity may leave the green roof more vulnerable to environmental change or disturbance events (33), while also potentially reducing building benefits (increased thermal performance) achieved by green roofs (23). Conversely, it is possible that the establishment of *S. scoparium*, like other pioneering grasses and cryptogams, may support the future emergence and colonization of stable, "higher-level" plant species and communities (36). It is therefore possible that, similar to the sedum species on House 2, the establishment of *S. scoparium* may eventually support the establishment of ruderal species and improve overall system diversity and resilience.

These changes in plant community raise questions about the role of emergent species in roof resilience and performance. If climate-adapted ruderal species positively contribute to biodiversity and increase the temporal and spatial stability of vegetative communities (16,27,33), can their presence be seen as performing meaningful service functions? Can we see value in these climate-adapted, resilient species, as we do with other plant diversity (23)?

While plant communities will continue to change and evolve over time, designers should be concerned with ensuring that a green roof is able to meet baseline performance goals over its lifetime. From this perspective, the ideal green roof should increase in resilience and value as it matures, proposing a vision of green roof design in which change is an essential component. In a context marked by variation and flux, stability and adherence to initial design conditions may not be the ultimate measure of success, and we may begin to rethink maintenance regimes imposed on these systems. Ultimately, long-term study of actual green roofs, integrated with building and site modeling as well as traditional forms of performance monitoring, will greatly improve both our understanding of green roof function over expected lifespans, as well as the design of future green roof systems.



Acknowledgments

The authors would like to thank Cornell University, particularly Art Fives and the members of Cornell's Offices of Construction Management, Grounds and Maintenance for assisting in this research and allowing access to their facilities.

References

- Banting, D., Doshi, H., Li, J., Missios, P., Currie, B. A., and Verrati, M., 2005. Report of the Environmental Benefits and Costs of Green Roof Technology for the City of Toronto. [pdf] Toronto: Ryerson University. Available at : http://www.toronto.ca/greenroofs/pdf/fullreport103105.pdf> [Accessed 01 June 2013].
- 2. Bass, B., 2009. *Biodiversity Research on Green Roofs: Developing a Research Protocol* [pdf]. Available at http://www.greenroofs.org/resources [Accessed 1 June 2013]
- Brenneisen, S., 2003. The benefits of biodiversity from green roofs: key design consequences. In: Cities Alive, *Proceedings of the 1st North American Green Roof Conference*. Chicago, 29-30 May 2003.
- 4. Butler, C. and Orians, C. M., 2009. Sedum facilitates the growth of neighboring plants on a green roof under water limited conditions. *Proceedings of the 7th North American Green Roof Conference: Greening Rooftops for Sustainable Communities.* Atlanta, GA 3-5 June 2009.
- 5. Butler, C. and Orians, C. M., 2011. Sedum cools soil and can improve neighboring plant performance during water deficit on a green roof. *Ecological Engineering*, 37(11), pp.1796-1803.
- Coffman, R., 2007. Comparing wildlife habitat and biodiversity across green roof type. Proceeds of the Fifth Annual Greening Rooftops for Sustainable Communities Conference, Awards and Trade Show. Minneapolis, MN 29 April-2 May November 2007. Toronto: Green Roofs for Healthy Cities.
- Cook-Patton, S. C. and Bauerle, T. L., 2012. Potential benefits of plant diversity on vegetated roofs: A literature review. *Journal of environmental management*, 106, pp.85-92.
- 8. Cornell Department of Horticulture, 2012. *Gardening Home*. [online] Available at: http://www.gardening.cornell.edu/weather/frezfree.html [Accessed 3 October 2012].
- 9. Dewey, D., Johnson, P. and Kjelgren, R., 2004. Species composition changes in a rooftop grass and wildflower meadow. *Native Plants.* 5, pp.56–65.



- 10. Dunnett, N. and Kingsbury, N., 2004. *Planting green roofs and living walls.* Vol. 254. Portland, OR: Timber Press.
- 11. Dunnett, N. and Hitchmough, J. eds., 2004. *The Dynamic Landscape: Design, Ecology and Management of Naturalistic Urban Planting.* London: Spon Press.
- Dunnett, N., Nagase, A., Booth, R. and Grime, P., 2008. Influence of vegetation composition on runoff in two simulated green roof experiments. *Urban Ecosystems*, 11(4), pp.385-398.
- Durhman, A.K., Rowe, D.B. and Rugh, C.L., 2007. Effect of substrate depth on initial growth, coverage, and survival of 25 succulent green roof plant taxa. *Hortscience*, 42, pp.588-595.
- 14. Franco, A. C. and Nobel, P. S., 1989. Effect of nurse plants on the microhabitat and growth of cacti. *The Journal of Ecology*, 77(3), pp.870-886.
- Getter, K. L., Bradley Rowe, D. and Cregg, B. M., 2009. Solar radiation intensity influences extensive green roof plant communities. *Urban Forestry & Urban Greening*, 8(4), pp.269-281.
- 16. Grime, J. P., 2002. *Plant strategies, vegetation processes, and ecosystem properties.* 2nd ed. West Sussex, England: John Wiley & Sons.
- Gunther, B., Larson, M. G. and Watt, F., 2010. LID and Sustainable Natural Resource Management in the Urban Environment: The Unique Case of New York City. In: Struck, S. and Lichten, K., ed. *Low Impact Development: Redefining Water in the City.* San Francisco: ASCE. 778-787.
- 18. Hansen, R. and Stahl, F., 1993. *Perennials and Their Garden Habitats*. Cambridge: Cambridge University Press.
- 19. Hill, M. O., 1973. Diversity and evenness: a unifying notation and its consequences. *Ecology*. 54, pp. 427–432.
- 20. Kadas, G., 2006. Rare invertebrates colonizing green roofs in London. *Urban Habitats*, 4, pp.66-86
- Kircher, W., 2002. Annuals and sedum-cuttings in seed-mixtures for extensive roof gardens. In: International Conference on Urban Horticulture 643, [online] Available at:<u>http://actahort.org/books/63/643_39.htm</u> [Accessed 1 June 2013].
- 22. Klinka, K., Chen, H. Y. H., Wang, Q. and de Montigny, L., 1996. Forest canopies and their influence on understory vegetation in early-seral stands on West Vancouver Island. *Northwest Science*, 70, pp.193-200.



- 23. Kolb, W. and Schwarz, T., 1993. Zum Klimatisierungseffekt von Pflanzenbeständen auf Dächern. *Veitshöchheimer Berichte*, 4, pp.28-36.
- 24. Köhler, M. and Poll, P. H., 2010. Long-term performance of selected old Berlin greenroofs in comparison to younger extensive greenroofs in Berlin. *Ecological Engineering*, 36(5), pp.722-729.
- 25. Lavorel, S. and Garnier, E., 2002. Predicting changes in community composition and ecosystem functioning from plant traits: revisiting the Holy Grail. *Functional Ecology*, 16(5), pp.545-556.
- 26. Lazzarin, R.M., Castellotti, F. and Busato, F., 2005. Experimental measurements and numerical modeling of a green roof. *Energy and Buildings*, 37, pp.1260-1267.
- Lundholm, J., MacIvor, J. S., MacDougall, Z., and Ranalli, M., 2010. Plant species and functional group combinations affect green roof ecosystem functions. *Plos One*, 5(3), p.9677.
- 28. Martin, M., 2007. *Native Plant Performance on a Seattle Green Roof*. Master's Thesis. University of Washington.
- 29. Monterusso, M. A., Rowe, D. B., and Rugh, C. L., 2005. Establishment and persistence of Sedum spp. and native taxa for green roof applications. *Horticultural Science*, 40(2), pp.391-396.
- 30. Mueller-Dombois, D & Ellenberg, H., 1974. *Aims and methods of vegetation ecology*. New York: Wiley Press.
- Northeast Regional Climate Center, Cornell University, 2013. Northeast Regional Climate Center. [online] Available at: http://www.nrcc.cornell.edu/> [Accessed 28 May 2013].
- NYC Department of Environmental Protection, 2012. NYC Green Infrastructure: 2012 Annual Report. [pdf] New York: NYC Department of Environmental Protection. Available at: http://www.nyc.gov/html/dep/pdf/green_infrastructure/gi_annual_report_2013.pdf> [Accessed 01 June 2013].
- Oberndorfer, E., Lundholm, J., Bass, B., Coffman, R., Doshi, H., Dunnett, N., Gaffin, S., Kohler, M., Liu, K., and Rowe, B., 2007. Green Roofs as Urban Ecosystems: Ecological Structures, Functions, and Services. *BioScience*, 57(10), pp.823 – 833.
- Rowe, Brad, 2011. Importance of Long-term Plant Evaluations for Extensive Green Roofs. In: *Cities Alive, Proceedings of the 9th North American Green Roof Conference*. Chicago, 30 November – 3 December 2011.



- 35. Shannon, C. E., 1948. A mathematical theory of communication. *The Bell System Technical Journal*, 27, pp.379-423, 623-656.
- Sutton, R. K., Harrington, J.A., Skabelund L., MacDonagh P., Coffman R.R., and Koch G. 2012. Prairie-based green roofs: literature, templates, and analogs. *Journal of Green Building* 7(1), pp.143-172.
- 37. Talbot, S. S., and Talbot, S. L., 1994. Numerical classification of the coastal vegetation of Attu Island, Aleutian Islands, Alaska. *Journal of Vegetation Science*, 5, pp.867–876.
- 38. Team, R.C., 2008. *R: A language and environment for statistical computing.* Vienna, Austria: R Foundation for Statistical Computing. *ISBN* 3-900051-07-0.
- 39. Theodosiou, T. G., 2003. Summer period analysis of the performance of a planted roof as a passive cooling technique. *Energy and Buildings*, 35(9), pp.909-917.
- 40. Turner, R.M. et al., 1966. The influence of shade, soil, and water on saguaro seedling establishment. *Botanical Gazette*, 127, pp.95-102.
- United States Department of Agriculture (USDA), 2012. Plant Hardiness Zone Map. [online] Available at: ">http://planthardiness.ars.usda.gov/PHZMWeb/Default.aspx">http://planthardiness.ars.usda.gov/PHZMWeb/Default.aspx">http://planthardiness.ars.usda.gov/PHZMWeb/Default.aspx
- 42. United States Department of Energy (USDE), YYYY. *Climate data based on EPW/STAT files from Elmira Regional Airport, NY located 35 miles to the Southwest*. [online] Available at: http://apps1.eere.energy.gov/buildings/energyplus/cfm/weather_data.cfm [Accessed 30 May 2013].
- 43. Villarreal, E. L. and Bengtsson, L., 2005. Response of a *Sedum* green-roof to individual rain events. *Ecological Engineering*, 25(1), 1-7.
- 44. Walker, M. D., Walker, D. A. and Auerbach, N. A., 1994. Plant communities of a tussock tundra landscape in the Brooks Range Foothills, Alaska. *International Journal of Vegetable Science*, 5, pp.843–66.
- Westhoff, V. and van der Maarel, E., 1978. The Braun–Blanquet approach. In: R.H. Whittaker, ed. 1980. *Classification of Plant Communities*. Houten, Netherlands: Springer Netherlands, pp.287–399.
- 46. Wolf, D., & Lundholm, J. T., 2008. Water uptake in green roof microcosms: Effects of plant species and water availability. *Ecological Engineering*, 33(2), 179-186.



Appendix

Latin Name	Common Name	Plant Type	Initial Cover	2012 % Cover	2012 Sociability
Schizachyrium scoparium	Little bluestem	Grass	23.23	65.00	66.71
Melilotus officinalis	Sweet clover	Forb	-	16.28	28.22
Panicum virgatum	Switchgrass	Grass	21.44	4.94	20.26
Festuca rubra	Creeping red fescue	Grass	5.00	4.04	13.70
Erigeron annuus	Fleabane	Forb	-	3.59	11.67
Liatris aspera	Rough blazing star	Forb	1.73	3.11	7.97
Lotus corniculatus	Birdfoot trefoil	Forb	-	2.48	21.39
Vicia spp.	Vetch	Forb	-	2.28	44.38
Hypericum perforatum	St. John's Wort	Forb	-	1.99	7.05
Bryophytes	Moss	Bryophyte	-	1.63	7.97
Lactuca serriola	Prickly lettuce	Forb	-	1.57	6.13
Phlox spp.	Phlox	Forb	-	1.38	26.89
Echinacea purpurea	Purple and white coneflower	Forb	2.42	1.19	11.56
Trifolium pratense	Red clover	Forb	-	1.19	7.71
Helianthus mollis	Downy sunflower	Forb	1.25	1.06	20.63
Asclepias tuberosa	Butterflyweed	Forb	1.73	1.06	10.31
Aster divaricatus	White wood aster	Forb	1.73	0.38	15.00
Heliopsis helianthoides	False Sunflower	Forb	1.25	0.26	5.63
Arctostaphylos uva-ursi	Bearberry	Shrub	22.75	0.22	8.75
Festuca ovina	Blue fescue	Grass	5.00	0.22	8.75
Eupatorim hyssopifolium	Hyssop leaved thoroughwort	Forb	1.25	0.19	15.00
Geranium maculatum	Wild geranium	Forb	2.20	0.03	2.50
Lupinus perennis	Wild blue lupine	Forb	2.20	-	-
Elymis hystrix	Bottlebrush grass	Grass	6.25	-	-

Latin Name	Common Name	Plant Type	Initial Cover	2012 % Cover	2012 Sociability
Sedum sexangulare	Tasteless stonecrop	Succulent	40.00	41.84	54.56
Sedum spurium 'Fuldaglut'	Two-row stonecrop	Succulent	32.00	32.80	55.22
Sporobolus heterolepsis	Prairie dropseed	Grass	20.00	7.80	37.05
Aster pilosis	Skinny aster	Forb	-	6.45	16.49
Melilotus officinalis	Sweet clover	Forb	-	6.26	13.88
Erigeron annuus	Fleabane	Forb	-	6.22	17.24
Daucus carota	Wild carrot	Forb	-	4.06	16.36
Allium cernuum	Nodding wild onion	Forb	6.00	4.02	38.21
Solidago canadensis	Goldenrod	Forb	-	3.61	13.71
Hypericum perforatum	St. John's wort	Forb	-	3.21	9.94
Setarius viridis	Foxtail	Grass	-	2.78	16.01
Trifolium pratense	Red Clover	Forb	-	2.61	15.11
Festuca rubra	Red fescue	Grass	-	2.58	10.38
Medicago lupulina	Medic	Forb	-	2.5	11.08
Taraxacum officinale	Dandelion	Forb	-	2.16	11.50
Populus species	Poplar species	Tree	-	1.80	10.00
Leontodon autumnalis	Fall dandelion	Forb	-	1.82	7.35
Solidago graminifolia	Lanceleaf goldenrod	Forb	-	1.56	10.92
Parthenocissus quinquefolia	Virginia creeper	Vine	-	1.47	10.26
Sempervivum species	Houseleeks	Succulent	2.00	0.06	2.50

Figure 6 House 1 (top) and House 2 (bottom) summary of initially planted species and all species identified in 2012 survey that were present on more than 1% of total green roof area. Species are ranked in order of dominance (percent cover) at time of 2012 census. Species in bold are from original planting in 2005 (Note: All originally planted species may not have been found to demonstrate greater than 1% cover in 2012).